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$[Mo_3S_{13}]^{2-}$ modified TiO_2 coating on non-woven fabric for efficient photocatalytic mineralization of acetone



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$$\label{eq:keywords:} \begin{split} \text{TiO}_2 \\ [\text{Mo}_3\text{S}_{13}]^2{}^- \\ \text{Photocatalytic oxidation} \\ \text{Non-woven fabrics} \\ \text{VOCs degradation} \end{split}$$

ABSTRACT

Improving the photocatalytic efficiency of commercial TiO_2 has important significance for practical application of TiO_2 based photocatalysts. A novel photocatalyst $[Mo_3S_{13}]^{2-}/TiO_2$ was fabricated by combining $[Mo_3S_{13}]^{2-}$ with commercial TiO_2 by an impregnation method. This composite photocatalyst presented a remarkable enhancement on photocatalytic mineralization of acetone in comparison with commercial TiO_2 . The optimum loading amount of $[Mo_3S_{13}]^{2-}$ was 1.7wt%, which is more efficient than the Pt/ TiO_2 (1.5wt%). Electrochemical impedance spectroscopy (EIS) showed the smooth electron transfer pathway in $[Mo_3S_{13}]^{2-}/TiO_2$ composite, facilitating the photo-charges separation during the photocatalysis process. Reactive oxygen species scavenging test illustrated that superoxide radical $(^{\circ}O_2^{-})$, hydroxyl radical $(^{\circ}OH)$ and photo-induced hole (h^+) were all contributing to the acetone degradation. The $[Mo_3S_{13}]^{2-}/TiO_2$ photocatalyst was deposited on non-woven fabrics which showed obvious promotion on the photocatalytic degradation of acetone in comparison with pristine commercial TiO_2 .

1. Introduction

People spend about 80% of their life indoors and the United States Environmental Protection Agency reported that indoor air is two to five more polluted than outdoor air [1,2]. Volatile and semi-volatile organic compounds are the dominating hazardous pollutants in the indoor and enclosed space, such as households, industrial areas, laboratories and office building [3-5]. As a green and environmental-friendly method, semiconductor photocatalytic oxidation has been widely investigated [6-12]. TiO_2 as one of the photocatalysts draw a mass of attention due to its excellent photocatalytic efficiency, high stability and low cost in degrading organic pollutants [5,10,11].

Whereas, the limited utilization of photo-induced carriers, electrons (e $^-$) and holes (h $^+$), suspends the practical application of ${\rm TiO_2}$. The conduction band (CB) e $^-$ transfer to oxygen molecule existing ubiquitously in the photocatalysis systems has been considered as the rate-determining step in semiconductor photocatalysis [13,14]. The slow interfacial electrons transfer was the primary reason for the less utilization of photo-induced charges [15,16]. As co-catalysts can capture and transfer the electrons efficiently and improve the charge

separation, co-catalysts are extensively used to ameliorate the photocatalytic activities of photocatalysts [14,8-18]. Pt, Ag and other noble metal were reported massively as the co-catalysts to improve the photocatalytic efficiency [19-21]. However, the high cost and the scarcity limited the large-scale application of noble-metal co-catalysts. Thus, non-noble metal as the co-catalysts drawn numerous attentions. As one of the most widely investigated non-noble metal co-catalysts, MoS₂ has shown an excellent performance in photocatalysis and electrocatalysis [22-25]. Besides of MoS₂, other molybdenum sulfides have also been identified as promising co-catalysts. Among them, thiomolybdate [Mo₃S₁₃]²⁻ contains three different types of sulfur ligands, which all resemble the active edge site of MoS2. It has been reported that the unique molecule structure of $[Mo_3S_{13}]^{2-}$ prominently facilitated the effective H₂ evolution in the hydrogen-evolution reaction (HER) [26-29]. However, to the best of our knowledge, there is no report about [Mo₃S₁₃]²⁻ modifying TiO₂ to study the photocatalytic properties in VOCs mineralization. In this paper, we used acetone as the model VOCs which is abundant in the cabins [30,31]. Our results indicated that [Mo₃S₁₃]²⁻ displayed excellent performance doubling the photocatalytic activity compared to the commercial TiO2 (P25). The

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predominant oxidant species during the photocatalysis process were also detected and discussed. Additionally, the photocatalysts were deposited on polypropylene non-woven fabric which is lightweight, antibacterial and low cost [32–34]. The [Mo₃S₁₃]² $^-$ /TiO₂ deposited non-woven fabric maintained the same increment on photocatalytic mineralization activity in comparison with pristine TiO₂/fabric.

2. Experimental

2.1. Chemicals

 ${
m TiO_2}$ powders (Degussa P25, ca. 80% anatase) were used as received. Non-woven 100% polypropylene fabric (65.3 g/m²) in white color was used as the substrate. Ammonium heptamolybdate tetrahydrate ((NH₄)₆Mo₇O₂₄·4H₂O), carbon disulfide, silicon dioxide, chloroplatinic acid hexahydrate, ethanol, acetone, *p*-benzoquinone, and toluene was of analytical purity and used as received. All of them were purchased from Sinopharm Chemical Reagent Co., Ltd. Ammonium polysulfide solution was purchased from Xiya reagent. Barium sulfate was purchased from Aladdin Industrial Corporation. Potassium iodide (KI) was purchased from Shanghai Macklin Biochemical Co., Ltd. Coumarin was purchased from J&K Scientific Ltd.

2.2. Materials synthesis

2.2.1. Preparation of $[Mo_3S_{13}]^{2-}/TiO_2$ and Pt/TiO_2

The thiomolybdate $(NH_4)_2Mo_3S_{13}$:n H_2O (n = 0–2) was prepared according to the literature [29]. The $[Mo_3S_{13}]^2$ was immobilized on TiO_2 by an impregnation method. In a typical process, different amounts of $[Mo_3S_{13}]^2$ solution was dispersed in the ethanol (30 mL, \geq 99.5%) to obtain a uniform mixture solution. Then 150 mg of TiO_2 powder (Degussa P25, ca. 80% anatase) were dispersed into the above mixture solution by sonification. The weight contents of $[Mo_3S_{13}]^2$ in the composites photocatalysts were adjusted from 0.6 wt%, 1.1 wt%, 1.7 wt% to 3.4 wt%. The suspension was stirred for 4.5 h before being centrifuged and washed. The $[Mo_3S_{13}]^2$ -/ TiO_2 composites were washed by affluent amounts of deionized water and ethanol before drying in a vacuum oven at 40 °C overnight. The $[Mo_3S_{13}]^2$ -/ TiO_2 composites were designated as $0.6[Mo_3S_{13}]^2$ -/ TiO_2 , $1.1[Mo_3S_{13}]^2$ -/ TiO_2 , $1.7[Mo_3S_{13}]^2$ -/ TiO_2 and $3.4[Mo_3S_{13}]^2$ -/ TiO_2 . For comparison, Pt/ TiO_2 photocatalysts with Pt of 1.5 wt% were synthesized according to the previous report [35].

2.2.2. $[{\rm Mo_3S_{13}}]^{2-}/{\rm TiO_2}$ photocatalysts coating on polypropylene non-woven fabrics

Different amounts (30 mg, 40 mg, 50 mg) of $[Mo_3S_{13}]^2$ - $/TiO_2$ photocatalysts were dissolved in 5 mL ethanol to obtain uniform suspensions. The obtained suspensions kept uniformly before being sprayed onto non-woven polypropylene fabric (area = $36cm^2$ and $0.25 m^2$) via an airbrush (U-STAR S-130) (Fig. S1). The supply gas pressure was controlled as 0.1 MPa. The treated fabrics were dried in a vacuum oven at $60\,^{\circ}\text{C}$ overnight. For comparison, the pristine TiO_2 particle was also loaded on the fabrics in the same way.

2.3. Characterization

X-ray diffraction analysis of the as-prepared samples were carried out on Shimadzu XRD-6100 diffractometer with Cu K α radiation ($\lambda=1.54\,\text{Å}$). The data were recorded at a scan rate of 5°/min in the 2θ range from 10° to 80°. UV–vis diffuse reflectance spectra (UV–vis DRS) were measured on UV–vis spectrophotometer (Shimadzu UV-2450, Japan). The elemental composition of the composites was determined by inductively coupled plasma optical emission spectrometer (ICP-AES, iCAP6300, Thermo, USA). The morphologies of samples were characterized with field emission scanning electron microscope (FESEM) (FEI & Oxford Sirion 200) and Transmission electron microscopy (TEM)

(JEOL-JEM-2010, Japan) employing a 200 kV accelerating voltage. Xray photoelectron spectroscopy (XPS) (AXIS Ultra DLD Kratos, Japan) analysis was carried out to study the surface electronic states of the samples. All binding energies were calibrated by the C 1s peaks of the surface adventitious carbon (284.6 eV). Fourier transform infrared (FTIR) spectra of the samples were measured on Tensor 27 FTIR spectrometer (Nicolet 6700). The surface physical adsorption property was studied through the contact angle (CA), which were measured using the sessile droplet method with a contact angle meter (KRUSS GmbH, Germany). The value of the contact angle was the average of five measurements on a sample. The electrochemical measurement was conducted in a three-electrodes system, where Pt plate as the counter electrode. Ag/AgCl electrode as the reference electrode and 1.0 M perchloric acid as the electrolyte, on ZAHNER PP211 electrochemical workstation. The working electrode was prepared by spreading 5 µL slurry on fluorine tin oxide (FTO) glass. The slurry contained 0.2 mL deionized water, $0.8\,\text{mL}$ ethanol $100\,\mu\text{L}$ Nafion and $2\,\text{mg}$ sample. The obtained glass was dried at room temperature overnight.

2.4. Photocatalytic activity evaluation

In order to evaluate the photocatalytic activity of as-synthesized $[{\rm Mo_3S_{13}}]^{2-}/{\rm TiO_2}$ composites, photocatalytic degradation of acetone was performed under a 300 W Xe arc lamp (21 A imported current, through a 50 \times 50 mm aperture winder). A 20 mg of as-prepared photocatalysts and 480 mg of SiO_2 were milled homogeneously before spreading evenly on a glass fiber plate (area of mixed powder, 16 cm²). In reactive oxygen species experiment, 2 mg scavenger was added into the mixture of photocatalysts and SiO_2 before milling. The vessel was pretreated by O_2 for 30 min to create an oxygen-rich environment. Then the acetone (6 μL) was injected into the vessel. After ensuring the sufficient mixture of gas in the closed vessel, the reactor was placed under the light irradiation. Every 30 min, a certain volume (500 μL) gas extracted from the vessel was injected into the gas chromatograph (GC-7900 equipped with a methane reforming furnace and flame ionization detector (FID)).

3. Results and discussion

The $[Mo_3S_{13}]^2$ -/TiO $_2$ composites with different mass ratios were synthesized through an impregnation method. The XRD patterns of pristine TiO $_2$ and $[Mo_3S_{13}]^2$ -/TiO $_2$ composites showed the typical characteristic diffraction peaks of TiO $_2$ containing anatase and rutile (Fig. 1). None of the newly appeared diffraction peaks assigning to $[Mo_3S_{13}]^2$ - were found on $[Mo_3S_{13}]^2$ -/TiO $_2$ nanocomposites indicating that the high dispersed loading of $[Mo_3S_{13}]^2$ - [6].

TEM analysis was carried out to investigate the crystal structure and the composition of the composites $(1.7[Mo_3S_{13}]^2-/TiO_2)$. As shown in

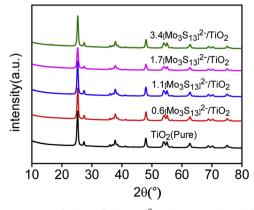


Fig. 1. XRD patterns of TiO_2 and $[Mo_3S_{13}]^2-/TiO_2$ composites with different contents of $[Mo_3S_{13}]^2-$.

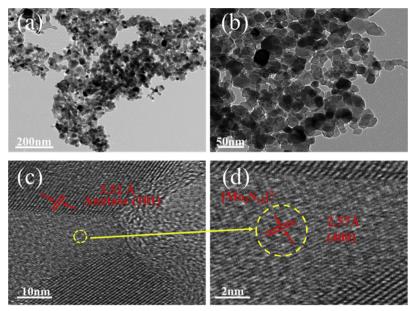


Fig. 2. TEM images and HRTEM images of $1.7[Mo_3S_{13}]^2$ /TiO₂ composites.

Fig. 2a and b, the P25 photocatalyst is composed of nanoparticles with the diameter of around 25 nm. High-resolution TEM (HRTEM) images were shown in Fig. 2c and d. The fringe spacing of 0.352 nm was consistent with the interplanar spacing (0.351 nm) of (101) lattice planes of anatase TiO₂ (Fig. 2(c)). The interplanar spacing of [Mo₃S₁₃]²⁻ was 0.257 nm (Fig. 2(d)), which is consistent with previous reports [28]. The results demonstrated the [Mo₃S₁₃]²⁻ had been successfully deposited on the TiO₂ nanoparticles surface by the impregnation method.

XPS spectrum of Ti 2p for pristine TiO2 emerged at 458.26 eV and 464.06 eV corresponded with Ti2p 3/2 and Ti2p 1/2, which illustrated the Ti⁴⁺ oxidation state of TiO₂ (Fig. 3a) [36,37]. Whereas, the peaks of binding energy of Ti 2p in the synthetic composites, 1.7[Mo₃S₁₃]²⁻/ TiO₂, altered slightly, locating at 458.33 eV and 463.93 eV, respectively. The spin-orbital splitting of Ti2p 3/2 and Ti2p 1/2 in the $1.7[Mo_3S_{13}]^{2-}/TiO_2$ was 5.6 eV, which was smaller than that of pristine TiO2 (5.8 eV). The results should be ascribed to the successful doping of $[Mo_3S_{13}]^{2-}$ on the TiO_2 and implied the interaction between $[Mo_3S_{13}]^{2-}$ and TiO_2 were not simply physical adsorption [5,6]. To confirm the existence of $[Mo_3S_{13}]^{2-}$, we also gave the XPS spectra of Mo 3d and S 2p. The characteristic peaks of Mo (Fig. 3(b)) were presented at 231.3 eV and 233.9 eV, which are attributed to the binding energy of Mo3d 5/2 and Mo3d 3/2 orbits with +4 oxidation state, respectively. This result was consistent with our previous research [6]. The S 2p spectrum (Fig. S2) was also consistent with previous reports [38,39]. FTIR spectrum (Fig. S3) of TiO₂ revealed three characteristic peaks at 3438, 1637, 659 cm⁻¹ assigned to stretching vibration of hydroxyl groups, bending vibration of surface O–H group and Ti–O–Ti bridging stretching mode, respectively [40,41]. Whereas, one new peak at $1384\,\mathrm{cm}^{-1}$ appeared in the FTIR spectrum of $[\mathrm{Mo_3S_{13}}]^2$ /TiO₂, which was associated to Mo–S bonding [39,42].

The gas-solid photocatalytic oxidation of acetone was studied over the $[Mo_3S_{13}]^{2-}/TiO_2$ composites and pristine TiO_2 by evaluating the CO₂ production under UV-vis light irradiation. Fig. 4(a) presented the CO2 production curve as a function of time. CO2 was not generated without the UV-vis light irradiation or the photocatalysts, which was demonstrated by the blank experiment in Figs. 4 and S4. In the presence of pristine TiO2, the CO2 production curve showed linear tendency and 49.3 µmol CO₂ was detected after 90 min irradiation. The deposition of [Mo₃S₁₃]²⁻ on TiO₂ improves the degradation of acetone and CO₂ production. The CO₂ production after 90 min irradiation was 65.4 µmol, 77.0 μ mol, 107.5 μ mol and 60.9 μ mol for 0.6 $[Mo_3S_{13}]^2$ /TiO₂, $1.1[\text{Mo}_3\text{S}_{13}]^2 - /\text{TiO}_2, \ \ 1.7[\text{Mo}_3\text{S}_{13}]^2 - /\text{TiO}_2 \ \ \text{and} \ \ 3.4[\text{Mo}_3\text{S}_{13}]^2 - /\text{TiO}_2$ composites, respectively. The results indicated that the optimal loading amount of $[Mo_3S_{13}]^{2-}$ is 1.7 wt%. It has been reported that $[Mo_3S_{13}]^2$ also greatly enhanced the electrocatalytic H₂ production for graphite paper, which indicates the electron capture capability of co-catalysts [Mo₃S₁₃]²⁻ [43]. Pt nanoparticles was an efficient co-catalyst in photocatalysis. Falch et al. reported that the optimal amount of Pt was 1.5 wt% in hydrogen production system [35]. Here we compared the CO₂ production process under the presence of 1.7[Mo₃S₁₃]²⁻/TiO₂ with that of $1.5\,\text{Pt/TiO}_2$. After 90 min irradiation, $84.4\,\mu\text{mol}$ CO_2 was detected which is less than that produced on 1.7[Mo₃S₁₃]²⁻/TiO₂. The above results supported that $[Mo_3S_{13}]^{2-}$ would become a satisfying

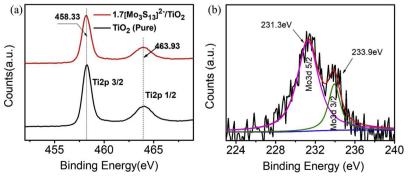


Fig. 3. XPS spectra of Ti 2p (a) and Mo 3d (b) of pure TiO_2 and the $1.7[Mo_3S_{13}]^{2-}/TiO_2$ composites.

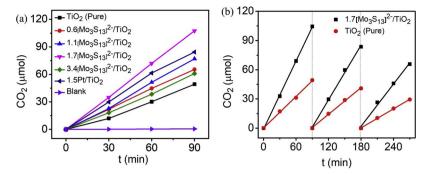


Fig. 4. (a) CO₂ production on different catalysts during the photocatalytic degrading of acetone; (b) CO₂ production in cycle runs of photocatalytic degradation of acetone.

alternative to noble metal co-catalyst, such as Pt. We also compared the photocatalytic degradation efficiency between [Mo₃S₁₃]²⁻/TiO₂ composite and some other reported photocatalysts. [Mo₃S₁₃]²⁻/TiO₂ composite simultaneously had high CO₂ initial production rate and high acetone mineralization ratio (Table S1). In addition, $[Mo_3S_{13}]^{2-}/TiO_2$ composite photocatalysts presented a hump-like variation tendency on photocatalytic activity relying on the loading amount of [Mo₃S₁₃]² The ascent stage as the increasing loading amount was ascribed to that [Mo₃S₁₃]²⁻ could facilitate the separation of photo-charges. The decline stage along with the increasing loading amount was due to that excess loading would shade the surface of TiO2, which suppressed the light absorption of photocarrier. The stability of composite photocatalysts was investigated by the cycling life test. As Fig. 4(b) shown, the initial reaction rates of the three cycles on the $1.7[Mo_3S_{13}]^{2-}/TiO_2$ test were 1.16 µmol/min, 0.94 µmol/min, 0.72 µmol/min, respectively. The initial reaction rate fell down about 37.9% in the third run. We also measured the initial reaction rates of the three cycles on the $1.1[Mo_3S_{13}]^{2-}/TiO_2$. The initial reaction rate fell down about 34.5% in the third run (Fig. S5). The cycling stability of the pristine TiO2 was measured for comparison. The initial reaction rates were 0.58 µmol/ min, $0.49 \,\mu\text{mol/min}$, $0.35 \,\mu\text{mol/min}$, respectively. The initial reaction rate fell down about 39.6% in the third run. The decay rates of initial reaction rate between [Mo₃S₁₃]²⁻/TiO₂ composites and pristine TiO₂ were almost the same. Our previous work indicated $[Mo_3S_{13}]^{2-}$ nanoclusters were intrinsically stable under the UV-vis irradiation [44]. Thus, the similar decay rates of initial reaction rate between [Mo₃S₁₃]²⁻/TiO₂ composites and pristine TiO₂ should be ascribed to the accumulation of intermediate product on TiO₂ [45,46].

The optical properties of pristine TiO_2 and the $[Mo_3S_{13}]^2$ – $/TiO_2$ composites were studied by UV–vis diffuse reflectance spectra (DRS). The light absorptions of photocatalysts increased obviously due to the coating of $[Mo_3S_{13}]^2$ – clusters with brown color (Fig. S6). To figure out the reason for the improved photocatalytic activity, the electrochemical impedance spectroscopy (EIS) was carried out, which is usually used to understand the internal electron transfer process [9,47]. The diameter of the semicircle in correlation with the charge-transfer resistance of the electrode is smaller for $[Mo_3S_{13}]^{2-}/TiO_2$ composite than that of

pristine TiO_2 in the high-frequency region (Fig. 5(a)), which should be attributed to $[Mo_3S_{13}]^{2-}/TiO_2$ composite with good electrical conductivity indicating the smooth electron transfer pathway in $[Mo_3S_{13}]^{2-}/TiO_2$ composite. Accordingly, the photo-induced charges separation efficiency was improved exceedingly, which benefited to the photocatalytic oxidation process. In addition, the polarization curves (Fig. S7) also indicated the efficient electrons transfer in composites [27].

The reactive oxygen species are the key points in the photocatalytic oxidation degradation. To understand the reaction mechanism in the degradation process, p-benzoquinone and KI were added into the reaction system as the superoxide radical ('O2") scavenger and photoinduced hole (h⁺) scavenger, respectively. As the Fig. 5(b) shown, the production of CO₂ was completely depressed when p-benzoquinone was added, indicating that 'O2 was the vital reactive species in this degradation system. The photo-induced electrons in the conduction band were transferred fleetly through [Mo₃S₁₃]²⁻ and reacted with O₂ absorbed in the surface of $[Mo_3S_{13}]^{2-}/TiO_2$ composites to form O_2^- , which could further oxidize organic compounds. Photo-induced holes also participated in the oxidation degradation process, according to the depressed production of CO2 when KI was added. The holes (h+), which have high oxidation potential, could oxidize directly the adsorbed compounds [6]. Besides of this route to contribute in the degradation process, h^+ (E_{VB} = 2.69 V at pH 7) could also react with the OH reactive site (E°(H_2O/OH) = 2.29 V) (e.g. H-bonded OH, terminal Ti-OH and Ti-OH-Ti bridged) [11,48] on the TiO2 to form the higher oxidation potential species hydroxyl radical ('OH), which participated in the photo-degradation process [49,50]. Specific signal molecule, coumarin could react with 'OH to form 7-hydrocoumarin, a strongly luminescent compound [51]. The fluorescence emission at 396 nm increased illustrating the production of 7-hydrocoumarin (Fig. S8). This result provided the evidence that 'OH was formed and participated in the process of mineralization of acetone (Fig. 6).

When the vessel was filled with Ar, there is no generated CO_2 could be detected. This phenomenon confirmed that O_2 is indispensable in photocatalytic degradation of gaseous pollutants as the electron acceptor. In the dry anaerobic atmosphere (Relative Humidity < 10%),

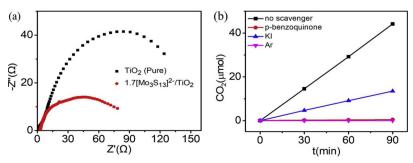


Fig. 5. (a) Nyquist plots of pristine TiO₂ and 1.7[Mo₃S_{13]}²⁻/TiO₂ composites; (b) CO₂ production with different reactive-species scavengers.

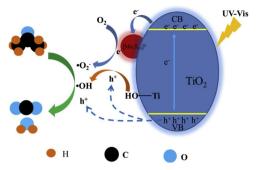


Fig. 6. The schematic illustration of the mineralization process of acetone under the presence of $[Mo_3S_{13}]^{2-}/TiO_2$ composite.

the photo-induced electrons in the conduction band of TiO_2 is hardly captured by intermediates such as O_2 , adsorbed water, and surface hydroxyl groups, and thus it would finally recombine with holes [15].

Polypropylene non-woven fabric is the promising supporter of powder photocatalysts due to their good resistance to alkalis and acids, low cost and recyclability [32]. We loaded pristine ${\rm TiO_2}$ and ${\rm [Mo_3S_{13}]^2}^-/{\rm TiO_2}$ composite on the fabrics through spraying method (Fig. S9). The morphologies of non-woven fabrics and non-woven fabrics coated with pristine ${\rm TiO_2}$ and ${\rm 1.7[Mo_3S_{13}]^2}^-/{\rm TiO_2}$ were investigated by SEM. Fig. 7(a) and (b) displayed a spatial three-dimensional structure constructed by plentiful reticular fibers, whose surface have numerous lousiness. The rough surface structure facilitated the depositing of photocatalysts. After deposition, the surface of fibers was homogeneously covered by the nanoparticles layers (Fig. 7(c) and Fig. 7(d)).

Water Contact Angle analysis were used to inquire into the surface character of the coated fabrics. The contact angle of pristine non-woven fabrics was 132.1° (Fig. 8(a)), which was ascribed to the hydrophobic nature of the fibers. On the contrary, the contact angle of the fabrics coated with pristine TiO_2 was 65.1° . This phenomenon was attributed to the hydrophilic properties of TiO_2 nanoparticles [52]. The fabric coated with $1.7[\text{Mo}_3\text{S}_{13}]^{2-}/\text{TiO}_2$ composite showed the contact angle at 117.8° , obviously bigger than that of fabrics coated with pristine TiO_2 . It has been reported that the $[\text{Mo}_3\text{S}_{13}]^{2-}$ contained the structural motif that resembles the edge site of MoS_2 , which has the hydrophobic characteristic [26,53]. Hence, this phenomenon also illustrated the

homogeneous distribution of $[Mo_3S_{13}]^{2-}$ on the TiO_2 surface.

The acetone photocatalytic degradation performance was investigated to evaluate the photocatalytic efficiency of non-woven fabrics coated with $[Mo_3S_{13}]^{2-}/TiO_2$. As shown in Fig. 9(a), CO $_2$ production presented a gradually rising and stable trend along with the increasing of loading amount of $1.7[Mo_3S_{13}]^{2-}/TiO_2$ on the fabrics. This phenomenon could be explained to that the more loading photocatalysts before it reaches the optimum loading amount (0.19 mg/cm 2 in this experiment) meant the more active sites exposed to the light irradiation. The excess photocatalysts exceeding the optimum loading amount cannot further improve the mineralization of acetone. We also compared the photocatalytic mineralization activity between fabrics coated with $[Mo_3S_{13}]^2-/TiO_2$ and fabrics coated with pristine TiO $_2$ (Fig. 9(b)). The CO $_2$ production rate for $[Mo_3S_{13}]^2-/TiO_2$ -fabric was $133.5\,\mu\text{mol}/(\text{min}\cdot\text{g})$, almost 3 times faster than that of TiO $_2$ -fabrics (43.2 $\mu\text{mol}/(\text{min}\cdot\text{g})$).

4. Conclusion

In summary, non-noble metal co-catalysts $[Mo_3S_{13}]^{2-}$ were successfully deposited on the commercial TiO_2 (P25). The photocatalytic mineralization of acetone illustrated that the $[Mo_3S_{13}]^{2-}/TiO_2$ showed greatly improved activity due to the efficient separation of photo-induced electrons and holes by $[Mo_3S_{13}]^{2-}$. $[Mo_3S_{13}]^{2-}$ is even more effective than Pt as co-catalyst on promoting the photocatalytic activity. The $[Mo_3S_{13}]^{2-}/TiO_2$ composite photocatalyst was successfully loaded on non-woven fabrics, which also showed enhanced photocatalytic mineralization activity than pristine TiO_2 modified fabric. Accordingly, non-woven fabrics coated with $[Mo_3S_{13}]^{2-}/TiO_2$ composites could be a promising photocatalytic material for indoor air purification.

Acknowledgments

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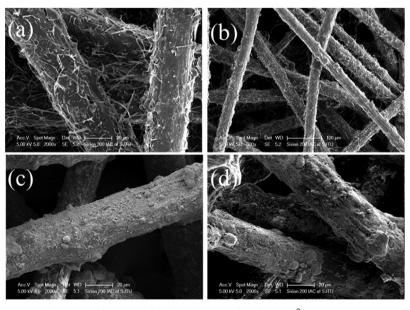


Fig. 7. SEM images of purified non-woven fabrics and fabrics coated with pure TiO_2 and $1.7[Mo_3S_{13}]^{2-}/TiO_2$ particles: (a) & (b) pure fabric, (c) TiO_2 -fabric, (d) $1.7[Mo_3S_{13}]^{2-}/TiO_2$ -fabric.



Fig. 8. Water droplets on different types of fabric surface: (a) pure fabric, (b) fabric coated with TiO_2 , (c) fabric coated with $1.7[Mo_3S_{13}]^2$ / TiO_2 composites.

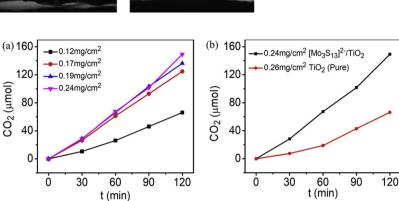


Fig. 9. (a) CO_2 production in the presence of fabrics coated with different amounts of $1.7[Mo_3S_{13}]^{2-}/TiO_2$ photocatalysts; (b) the comparison of CO_2 production in the presence of fabrics coated with TiO_2 and $1.7[Mo_3S_{13}]^{2-}/TiO_2$ photocatalysts.

Appendix A. Supplementary data

Supplementary material related to this article can be found, in the online version, at doi:https://doi.org/10.1016/j.apcatb.2018.12.060.

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